Cost of inaction

Summary

The present report was prepared by the Task Force on Integrated Assessment Modelling in cooperation with the Task Force on Techno-economic Issues in accordance with item 2.1.5 of the 2022–2023 workplan for the implementation of the Convention (ECE/EB.AIR/148/Add.1). The report aims to encourage ratification and implementation of the protocols to the Convention, in particular the Protocol to Abate Acidification, Eutrophication and Ground-level Ozone, as amended in 2012, demonstrating to policymakers the comparison of the costs of inaction on air pollution – defined as the damage to health, ecosystems and economy (mainly via work absenteeism) – with the costs of taking action, defined as the costs of abatement measures.

The present document is being presented to the Working Group on Strategies and Review for consideration. It is expected that a final draft will then be forwarded to the Executive Body for adoption at its forty-second session (Geneva, 12–16 December 2022, tentatively).
I. Key messages

1. In nearly half of the countries in the United Nations Economic Commission for Europe (ECE) region (26 of 56) the current monetary damage costs to health and ecosystems due to ambient air pollution corresponds to over 5 per cent of gross domestic product (GDP). In at least 6 countries, the damage is more than 10 per cent of GDP. The largest part of the damage costs consists of reduced life expectancy, followed by morbidity costs (e.g., hospital admittance, sick leave, medicine costs), and damage to ecosystems. The monetized damage is – as a percentage of GDP – higher in the Eastern than in the Western part of the ECE region. Globally, labour productivity losses (mainly via work absenteeism) due to air pollution make up approximately 5–9 per cent of the total damage costs.

2. There are societal values yet to be monetized and included in the damage cost estimates, foremost the damage to biodiversity. There are also considerable information gaps between the Eastern and Western parts of the ECE region, especially regarding valuation studies carried out by East European research groups and scenarios for future air pollution levels in Eastern Europe. Dedicated efforts are still needed to address these missing values and gaps.

3. Thanks to existing policies, monetary damage in the Cooperative Programme for Monitoring and Evaluation of the Long-range Transmission of Air Pollutants in Europe (EMEP) domain up to 2030 is expected to be reduced by at least 14 per cent compared to 2020. The implementation of national emission reduction obligations and current emission limit values for vehicles, installations, non-road mobile machinery and products will reduce damages. Expected damage reduction will (as a percentage of GDP) be higher in the Western part of the ECE region since this region is expected to implement stricter emission reductions.

4. Up to 21 per cent of monetary damage in the European Union-27 in 2030–2050 could be avoided by additional (not included in current legislation) policy actions targeting air pollution. Applying technically feasible measures (not entailing excessive costs) could reduce annual monetary damage by 4 per cent (compared to the baseline) in 2030–2050; further damage reduction (to 20–21 per cent) can be achieved by applying all possible air pollution measures regardless of costs (maximum technically feasible reduction (MTFR) scenario). If the MTFR scenario is combined with climate measures, the damage reduction in 2050 might reach 26 per cent. In the Eastern part of the ECE region in particular, there insignificant potential for reducing monetary damage.

5. Abatement costs (costs of taking action) are significantly lower than those of inaction. Benefits tend to be higher than costs. In the European Union-27, abatement costs of available additional actions on top of current ambition levels in National Air Pollution Control Programmes (NAPCPs) are over 20 times lower than costs of avoided damage.

II. Introduction

6. Since the 1960s, economists have developed methods to monetize welfare effects of adverse ecosystem and human health effects caused by poor air quality. Although early (and partly biased) attempts showed that costs of reducing emissions far exceeded the benefits, it is now well established that the situation is the opposite in almost all cases. Failure to act to improve air quality is therefore imposing avoidable welfare losses. In other words, failure to take action leads to costs of inaction.

7. In preparing this report, the best available knowledge on damage costs of air pollution has been reviewed and synthesized. Based on this review and synthesis of the state-of-the-art science in this discipline, the most important messages to policymakers have been extracted. The work has been guided by the following questions:

   (a) Can welfare effects of poor air quality be confidently estimated?
   (b) How high are the damage costs when action is not taken on air pollution?
   (c) Are these damage costs expected to go up or down in the future?
   (d) How can the costs of inaction be further reduced?
(e) Will human welfare improve if more is done?

8. Included below is a conceptual overview of the costs of inaction and the most relevant literature. This is followed by an estimate of the current amount of damage costs from air pollution in the ECE region, the expected future reduction in damage costs, as well as available improvement potential.

III. Monetizing damages from air pollution

9. Although the exact terminology differs from one practitioner to another, in this report the following terminology was considered when writing about the economic effects of poor air quality. Welfare losses for society of poor air quality come in two main types: market costs and non-market costs (see figure 1 below).

Figure 1
Market and non-market costs of air pollution damage, split into their main categories. Figure copied from OECD, 2016.1

10. As examples of market costs, it has been shown that poor air quality causes productivity losses by reducing both the number of days people can go to work and forest, crop and vegetable harvest sizes. Furthermore, mitigating the negative effects of poor air quality consumes societal resources, such as health-care sector expenditures. All these are directly measurable costs that depend on current market prices and labour and health-care costs. More indirect market costs also exist – for instance, the reduction of available financial resources for investments. Methods and data needed to estimate market costs are clearly described in recent Organisation for Economic Co-operation and Development (OECD) reports (Atkinson et al., 2018;2 OECD, 2016).3

11. Non-market costs occur because poor air quality reduces people’s quality of life through illness leading to pain, suffering and discomfort, and through preterm mortality. Non-market costs are the type of damage costs typically dominating the value of benefits in cost-benefit analyses supporting policy decisions. Non-market costs do not directly result in expenses and cannot be quantified in the same way as market costs. There is a range of studies setting economic values on mortality and morbidity based on individuals’ perception of the value of a change in life expectancy, risk of fatal accidents, or health status – i.e., by applying willingness-to-pay methods. To assess costs of premature mortality due to air pollution, two main approaches exist – one involves a valuation metric called the Value of Life Year

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1 OECD, 2016. The Economic Consequences of Outdoor Air Pollution – Policy Highlights. Note that the OECD did not include (non-market) ecosystem damage in their approach.
(VOLY), another uses the Value of Statistical Life (VSL) (see box 1 below). Whether VOLY or VSL is used in a mortality cost estimate can significantly affect the results – therefore, the authors of the present report try to indicate the chosen metric in the numbers presented,\(^4\) where possible by writing VSL or VOLY after the value presented.

**Box 1**

**Value of Life Year and Value of Statistical Life approaches to valuation of premature mortality caused by air pollution**

The VOLY and VSL approaches to valuation of life-shortening from air pollution differ since the life years lost from air pollution typically is around 11 years, which is lower than the halved life expectancy typically associated with VSL studies. In detail, the VOLY method is based on life tables; it takes into account at what age people die from air pollution and gives results in terms of life expectancy. The VSL method does not use life tables and instead operates with mortality rates. As the VSL method does not take into account age or death reasons, it is sometimes considered to be overestimating health benefits from air pollution reduction (Desaigues et al., 2011)\(^5\) while VOLY approach is considered as more conservative. On the other hand, the VOLY approach is criticized for not valuing vulnerable populations (sick and elderly) as high as average populations.

Operationally in most impact assessments, the effect of air pollution on life expectancy or mortality is calculated through a fixed % change on a baseline life expectancy or relative risk. An outcome of this method is that in countries with short baseline life expectancy (i.e., high relative risk of mortality), air pollution affects more life years. Correspondingly, the numerical difference between mortality valuation with the VSL and VOLY approach will be higher in countries with long life expectancy and lower in countries with short life expectancy.


12. In this report, the term “damage costs” refers to the sum of all the above-mentioned cost types and categories.

**IV. Data, method, sources**

13. This report summarizes the most recent knowledge about current and projected damage costs due to air pollution and the costs of taking action to reduce damage. Most of the data are found in relevant articles and reports published in the last 10 years. Additionally, the authors conducted supplementary analysis of region-specific health damage utilizing the widely used Greenhouse Gas and Air Pollution Interaction and Synergies (GAINS) and Alpha RiskPoll models.

**A. Modelling made in this report as input into the data synthesis**

14. In 2020, the International Institute for Applied System Analysis (IIASA) published scenarios for the second Clean Air Outlook\(^1\) exploring different future air quality levels of ambitions regarding air pollution and climate measures in the European Union-27. The underlying baseline GAINS scenario (CAO2_Baseline_2030 in the scenario group Clean Air Outlook 2) is publicly available and reflects current and projected development in the entire GAINS modelling domain. To estimate current and projected health damage in the selected regions, including countries outside the European Union-27 (described below), the authors

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\(^4\) Where both options were available, the authors of the present document chose to present the numbers in Value of Statistical Life (VSL) – i.e., when presenting the results of the authors’ own calculations based on Amann, M. et al., 2020, Support to the development of the Second Clean Air Outlook. VSL was chosen since this metric, unlike Value of Life Year (VOLY), allows equal valuation for lives of people of different ages and “pre-existing conditions”. Furthermore, VSL was used by OECD in its recent studies frequently referred to in this report. The authors of the present report are, however, aware that the European Commission more often uses VOLY in its assessments and policy suggestions.

\(^5\) Amann, M. et al., 2020. Support to the development of the Second Clean Air Outlook.
of the present report used the GAINS model outputs (population-weighted concentrations of fine particulate matter of less than 2.5 microns in diameter (PM$_{2.5}$) and Sum of Ozone Means Over 35 ppb (SOMO35) in the receptor countries) as inputs into the Alpha RiskPoll model where the health effects of and corresponding damage from PM$_{2.5}$ and ozone are calculated and aggregated by region. Damage assessment is done for 2020 (current situation) and 2030 (projection).

**B. Regionalization**

15. Within the assessment of the monetized damage from air pollution to human health and ecosystems in this report, total and unit damage costs are summarized and analysed separately for each of the following subregions of the ECE region:

(a) EECCA: Armenia, Azerbaijan, Belarus, Georgia, Kazakhstan, Kyrgyzstan, Republic of Moldova, Russian Federation, Tajikistan, Turkmenistan, Ukraine, Uzbekistan;

(b) South-Eastern Europe: Albania, Bosnia and Herzegovina, Montenegro, North Macedonia, Serbia, Turkey;

(c) Western and Central Europe: European Union-27, Iceland, Norway, Switzerland, United Kingdom of Great Britain and Northern Ireland;

(d) North America: Canada, United States of America.

**C. Currency recalculations and value transfer**

16. All monetary estimates in the present report are presented in year 2015 euros unless otherwise indicated. When translating non-market (intangible) health damage estimates available in the literature into year 2015 euros, total inflation is accounted for (Consumer Price Index (CPI)) and change in GDP per capita Purchasing Power Parity in the considered country or region and a VSL income elasticity of 0.8 (as recommended in OECD, 2012) is applied. For estimates of technical costs and damage costs from the literature that include a large share of market-based costs (e.g., crop damage costs, labour productivity cost assessments), values are recalculated to year 2015 euros considering CPI only.

17. In the damage estimates based on GAINS and Alpha RiskPoll made for this report, valuation of health effects is harmonized with the recommended values used in Amann et al., 2020. Health damage value estimates are first translated from year 2005 euros to year 2015 euros by applying CPI and change in GDP per capita PPP across the European Union-28 with an income elasticity of 0.8 (as in Amann et al., 2020). Then, depending on the type of results presented, one of the two spatial value transfer methods is chosen:

(a) When assessing damage as a percentage of a country’s GDP, the country-specific damage is adjusted with the income difference between the considered country and the European Union-27. An income elasticity of 0.8 is assumed for countries with an income higher than the European Union average, and one of 1.2 for countries with an income lower than the European Union average. The adjusted values are compared to GDP PPP;

(b) When presenting the absolute damage numbers per region (EECCA/South-Eastern Europe/Western and Central Europe) in year 2015 euros, the damage is adjusted with the income difference between the ECE (Europe) region and the European Union-27, applying an income elasticity of 1.2 (since average income in the ECE (Europe) region is lower than in the European Union-27).

18. It is worth noting that some of the morbidity-related costs (market costs) are estimated with other methods than willingness-to-pay, so adjusting all morbidity values with respect to

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7 A VSL income elasticity of 1.2 is recommended for lower- and middle-income countries and an elasticity of 0.8 for higher income countries (Narain, U., Sall, C. 2016. Methodology for valuing the health impacts of air pollution – Discussion of challenges and proposed solutions).
income elasticity leads to some underestimation of cost of inaction (COI). However, since over 90 per cent of health damage is attributable to mortality, as well as to pain and suffering from illness, the underestimation of applying a 0.8 income elasticity has an insignificant effect on total COI.

19. The estimates of the authors of the present report of the damage as a percentage of a country’s GDP are done for the year 2020 (for 2010 for some countries of the Caucasus and Central Asia). These values should not be mixed up with values referred to as “per cent of GDP change” adopted from an OECD study – those are percentages of GDP in 2060, compared to the baseline scenario.

20. All data used to convert literature values and Alpha RiskPoll values to year 2015 euros are taken from the World Bank, OECD and the European Central Bank. GDP PPP, GDP per capita PPP and population data were downloaded from the World Bank database. CPI was obtained from the OECD database. Currency exchange rates were taken from the European Central Bank website.

V. How large is the monetized damage from air pollution to human health and ecosystems?

21. The first question answered in this overview relates to the total size of the damage costs and is presented for the EECCA, South-Eastern Europe, Western and Central Europe, North America and Global regions. EECCA is split between the countries within the Cooperative Programme for Monitoring and Evaluation of the Long-range Transmission of Air Pollutants in Europe (EMEP) domain and the countries outside that domain.

22. Country-specific total damage from air pollution and unit damage of specific pollutants depend on various factors such as population density, age structure and health, current state of crops and forests, proximity to neighbouring countries and weather conditions affecting transboundary pollution. Depending on where emissions occur, the resulting damage in any considered region would usually be different.

A. Countries of Eastern Europe, the Caucasus and Central Asia

Total damage and per cent of gross domestic product

23. Total health damage from air pollution in the EECCA countries within the EMEP domain shows a descending trend (see figure 2 below, left-hand panel). Damage is expected to fall by €17.5 billion (4 per cent) between 2020 and 2030 – still, annual damage will account to €425 billion in 2030. Damage from air pollution for countries with national borders within the EMEP domain corresponds to 5–7 per cent of the countries’ GDP (see figure 2, right-hand panel). For the Russian Federation, damage is only assessed for the European part of the country. Compared to the entire country’s GDP, it constitutes 6 per cent; however, this value does not account for health damage to the 25 per cent of the country’s population living east of the Ural Mountains and is therefore an underestimation.

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9 https://stats.oecd.org/
12 Hereinafter, EMEP domain as represented in GAINS Europe (v.3).
24. For the Caucasus and Central Asia outside the EMEP domain, damage from preterm mortality attributable to poor air quality varied from approximately €9.8 billion in Armenia to over €70 billion in Uzbekistan in 2010 (see figure 3 below). This estimate is based on the mortality rates due to ambient air pollution presented in the World Health Organization (WHO) and OECD, 2015, and a VSL value of €3.06 million (year 2005 euros) as in Amann et al., 2020. Mortality-related damage attributable to air pollution corresponds to 3–12 per cent of the countries’ GDP.

Figure 3
Health damage costs from ambient air pollution in the Caucasus and Central Asia in 2010 (based on mortality presented in World Health Organization and Organisation for Economic Co-operation and Development, 2015)
Reduced labour productivity and other morbidity effects

25. Costs of reduced labour productivity (lost working days) due to illness constitute about 0.6 per cent of total health damage costs, while all morbidity effects correspond to 5 per cent of total damage. OECD, 2016 estimates that damage from morbidity is in all regions dominated by costs of restricted activity days – for example, in the Russian Federation, costs of illness in 2060 are projected to be about €205 per capita, of which 68 per cent is attributable to restricted activity. The effect of pollution on labour productivity is expected to reduce the GDP of the Russian Federation by 0.8 per cent in 2060.

Damage to crops

26. In addition to premature mortality and illness in the population, air pollution has negative effects on crops – mainly through plants’ exposure to ground-level ozone. These effects are easily monetized through market prices. OECD, 2016 projects that, in the Russian Federation, by 2060, crop yields will be 5 per cent lower than in the “no-feedback” projection, corresponding to a 0.8 per cent decrease in value added in the agricultural sector.

Costs by pollutants

27. Costs of air pollution per ton emissions of main pollutants are presented in table 1 below. The values are obtained from the detailed modelling of pollutant transfer and effects on health and environment (Schucht et al., 2021); they represent health damage in the emitter country caused by PM$_{2.5}$ and ozone precursors. These damage values can be easily applied in damage costs assessments supporting air quality-related decision-making.

Table 1

<table>
<thead>
<tr>
<th>Country</th>
<th>NO$_x$</th>
<th>PM$_{2.5}$</th>
<th>SO$_2$</th>
<th>NMVOCs</th>
<th>NH$_3$</th>
</tr>
</thead>
<tbody>
<tr>
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<td>311 800</td>
<td>73 800</td>
<td>7 000</td>
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<td>28 100</td>
<td>400</td>
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<td>20 400</td>
<td>100</td>
<td>11 000</td>
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<tr>
<td>Georgia</td>
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<td>448 600</td>
<td>68 800</td>
<td>3 500</td>
<td>16 400</td>
</tr>
<tr>
<td>Rep. of Moldova</td>
<td>7 000</td>
<td>105 200</td>
<td>17 900</td>
<td>100</td>
<td>19 800</td>
</tr>
<tr>
<td>Russian Federation (EMEP)</td>
<td>4 500</td>
<td>110 500</td>
<td>34 700</td>
<td>1 400</td>
<td>37 700</td>
</tr>
</tbody>
</table>

Note: Data for Ukraine are not presented in Schucht et al., 2021; and the figures from the previous modelling of damage from industrial air pollution in Europe (Holland et al., 2014) are not comparable to those presented in table 1 above due to methodological differences, thus values for Ukraine are not included.

Abbreviations: NH$_3$, ammonia; NMVOCs, non-methane volatile organic compounds; NO$_x$, nitrogen oxides; SO$_2$, sulfur dioxide.

15 “No-feedback” baseline projection describes hypothetical baseline developments in absence of feedback effects of air pollution on the economy.
B. South-Eastern Europe

Total damage and per cent of gross domestic product

28. Health damage from ambient air pollution in South-Eastern Europe shows an ascending trend (see figure 4 below, left-hand panel): it is expected to increase from €100 billion in 2020 to €116 billion in 2030 (17 per cent increase) if no additional action is taken. Health damage attributable to air pollution constitutes 4–15 per cent of the countries’ GDP in this region (see figure 4, right-hand panel).

Figure 4
Health damage costs from air pollution in South-Eastern Europe (own calculations based on current legislation scenario in Amann et al., 2020)\(^5\)

Reduced labour productivity and other morbidity effects

29. Costs of reduced labour productivity due to illness equal 0.7 per cent of total health damage costs, and all morbidity effects constitute 10 per cent of total health damage.

Costs by pollutants

30. Pollutant-specific unit damage costs for South-Eastern Europe are summarized in table 2 below. These costs represent health damage in the European Environment Agency (EEA)\(^38\)+United Kingdom of Great Britain and Northern Ireland caused by PM2.5 and ozone precursors emitted in the listed countries. If health impacts of nitrogen dioxide (NO\(_2\)) precursors, impacts on crops and forests and material damage were to be added to the estimates, damage costs would increase by 2 per cent for NMVOCs, and by 43 per cent for NO\(_x\) (on average in the considered region). As in EECCA, the greatest damage per ton pollutant results from emissions of PM\(_{2.5}\), and the smallest from NMVOCs.
Table 2
Damage costs from air pollutants in South-Eastern Europe, year 2015 euros/ton, Value of Statistical Life (source – Schucht et al., 2021, table 21).

<table>
<thead>
<tr>
<th>Country</th>
<th>NO&lt;sub&gt;x&lt;/sub&gt;</th>
<th>PM&lt;sub&gt;2.5&lt;/sub&gt;</th>
<th>SO&lt;sub&gt;2&lt;/sub&gt;</th>
<th>NMVOCs</th>
<th>NH&lt;sub&gt;3&lt;/sub&gt;</th>
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<tr>
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<td>148 900</td>
<td>46 000</td>
<td>1 900</td>
<td>21 800</td>
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<td>Bosnia and Herzegovina</td>
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<td>40 600</td>
<td>2 700</td>
<td>50 600</td>
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<td>1 700</td>
<td>30 700</td>
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<td>3 000</td>
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</tr>
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<td>1 700</td>
<td>23 400</td>
</tr>
</tbody>
</table>

C. Western and Central Europe

Total damage and per cent of gross domestic product

31. For all of Western and Central Europe, total health-related damage from air pollution is estimated at approximately €980 billion in 2020 (see figure 5 below, upper panel). By 2030, this figure is expected to decrease to approximately €760 billion (a reduction by 22 per cent). In relation to GDP, country-specific damage varies from 1 per cent to around 14 per cent (the average value is 5 per cent) (see figure 5, lower panel).

Figure 5
Health damage costs from air pollution in Western and Central Europe (own calculations based on the current legislation scenario in Amann et al., 2020).

Abbreviations: AT, Austria; BE, Belgium; BG, Bulgaria; CH, Switzerland; CY, Cyprus; CZ, Czechia; DE, Germany; DK, Denmark; EE, Estonia; ES, Spain; FI, Finland; FR, France; GB, United Kingdom of Great Britain and Northern Ireland; GR, Greece; HR, Croatia; HU, Hungary; IE, Ireland; IT, Italy; LT, Lithuania; LU, Luxembourg; LV, Latvia; MT, Malta; NL, Netherlands; NO, Norway; PL, Poland; PT, Portugal; RO, Romania; SE, Sweden; SI, Slovenia; SK, Slovakia.
Reduced labour productivity and other morbidity effects

32. Costs of lost working days constitute about 1.1 per cent of total health damage costs; all morbidity effects account for 7 per cent of total damage (for comparison – Holland et al., 2014\textsuperscript{18} estimated the share of morbidity in total health damage from air pollution at about 9 per cent). According to a recent study exploring air pollution damage in Finland (Kukkonen et al., 2020),\textsuperscript{19} productivity losses accounted for 0.3–3.4 per cent of health damage in 2015, depending on the emission source (the largest impact on labour productivity is observed for PM\textsubscript{2.5} emissions from non-road machinery in urban areas). Expected GDP reduction in 2060 due to labour productivity losses in Western and Central Europe is 0.1–0.3 per cent (OECD, 2016)\textsuperscript{1}.

33. Levels and sources of air pollution are different in rural areas and in cities. Damage from air pollution in European cities exceeds €150 billion and largely depends on transport policies and corresponding emissions (see box 2 below).

Box 2

\textit{Zooming in on air pollution – city-level perspective (source - CE Delft, 2020\textsuperscript{20}).}

Recent analysis of health-related damage from air pollution in 432 large Western European cities (CE Delft 2020\textsuperscript{20}) estimates the total damage at over €2015 156 billion in 2018. From this, 76% is attributable to mortality (VOLY) while 24% - to pain and suffering from illness. Annual damage per capita is €2015 1250, which corresponds to ~9% of the cities’ income. City size is identified as a key factor in the social costs of air pollution.

The study highlights the link between transport policies and the social costs of air pollution. It is estimated that a 1% increase in the number of cars in a city results in 0.5% increase of the air pollution-related damage.

Damage to crops

34. Production of crops and wood in Europe is reduced by up to 15 per cent due to the harmful effects of ground-level ozone, depending on species sensitivity; annual losses for wheat production are estimated to be over €46 billion (Maas and Grennfelt, 2016).\textsuperscript{21} A more recent study of effects of air pollution on crops and vegetables in France (Schucht et al., 2019,\textsuperscript{22} see figure 6 below) indicates that the monetized damage might be larger. The study estimates that, in France alone, current economic losses for production of crops and wood amount to approximately €2.4 billion. Damage is expected to decrease by 10 per cent within the next decade – still, it will be equivalent to 8 per cent of costs of health damage from air pollution in France.

\textsuperscript{18} Holland, M., 2014. Cost-benefit Analysis of Final Policy Scenarios for the EU Clean Air Package, corresponding to IIASA TSAP Report #11.


\textsuperscript{20} de Bruyn, S., de Vries, J., CE Delft, 2020. Health costs of air pollution in European cities and the linkage with transport.


35. In Europe, the main sector contributing to air pollution is transport (González Ortiz et al., 2020).\textsuperscript{23} Total annual damage costs from road transport in the European Union-28 are estimated at up to €80 billion (CE Delft, 2018),\textsuperscript{24} with large variations between countries illustrated in figure 7 below. About 75–83 per cent of damage from road transport is attributable to diesel sources.

**Figure 7**

Traffic-related air pollution damage costs in 2016 for nine European countries.\textsuperscript{24}

Note the logarithmic scale of Y-axis.

36. Available estimates of sector-specific costs of air pollution per ton emissions indicate large variations between sectors and pollutants (see figure 8 below).

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\textsuperscript{24} CE Delft, 2018. Health impacts and costs of diesel emissions in the EU.
Figure 8
Sector-specific damage cost ranges from air pollutants in Western and Central Europe, thousands year 2015 euros/ton

Abbreviations: AGR, agriculture; Road, road transport; IND/PP, industries and energy; DOM, residential combustion.

37. Costs of air pollution per ton emissions (see table 3 below) also vary between countries depending on factors such as population structure and proximity to other countries. These costs represent health damage in EEA38+United Kingdom of Great Britain and Northern Ireland caused by PM$_{2.5}$ and ozone precursors emitted in the listed countries. If health impacts of NO$_2$ precursors, impacts on crops and forests and material damage were to be added in the estimates, damage costs would increase by 1 per cent for SO$_2$, 3 per cent for NMVOCs, and 134 per cent for NO$_x$ (on average in the considered region). Unit costs of PM$_{2.5}$ are high (up to €540,000/ton), while unit costs of NMVOCs have the lowest values (€400–€14,000/ton), indicating the same relative input into total damage from different pollutants as in other countries in Europe.

25 Swedish Road Administration, 2018. Analysmetod och samhällsekonomiska kalkylvärden för transportsektorn: ASEK 6.1
Table 3
Damage costs from air pollutants in Western and Central Europe, year 2015 euros/ton, Value of Statistical Life (source – Schucht et al., 2021, table 21).

<table>
<thead>
<tr>
<th>Country</th>
<th>NO&lt;sub&gt;x&lt;/sub&gt;</th>
<th>PM&lt;sub&gt;2.5&lt;/sub&gt;</th>
<th>SO&lt;sub&gt;2&lt;/sub&gt;</th>
<th>NMVOCs</th>
<th>NH&lt;sub&gt;3&lt;/sub&gt;</th>
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</table>
38. CE Delft, 2018\textsuperscript{28} provides aggregated damage unit costs for the European Union-28 (see figure 9 below) that also show that the highest damage per ton emission occurs from PM\textsubscript{2.5}.

Figure 9

Damage costs from air pollutants in the European Union-28, high Value of Life Year

(source – CE Delft, 2018)\textsuperscript{28}.

39. Estimates of historical total annual damage from air pollution in the United States of America and Canada vary from €145 billion to over €1,000 billion (0.4–8 per cent of GDP), depending on year, effects considered, and chosen valuation metrics (see table 4 below).
Table 4
Estimates for damage from air pollution in North America, in billions of year 2015 euros

<table>
<thead>
<tr>
<th>Country</th>
<th>Year</th>
<th>Damage</th>
<th>Per cent of GDP</th>
<th>Included effects; chosen metric for valuation (if available)</th>
<th>Source</th>
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</thead>
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<tr>
<td>United States</td>
<td>2010</td>
<td>150</td>
<td>1</td>
<td>Mortality, morbidity; VOLY</td>
<td>Im et al., 2018</td>
</tr>
<tr>
<td>United States</td>
<td>2011</td>
<td>510</td>
<td>3</td>
<td>Mortality; VSL</td>
<td>Goodkind et al., 2019</td>
</tr>
<tr>
<td>United States</td>
<td>2014</td>
<td>340</td>
<td>2</td>
<td>AP3 IAM model</td>
<td>Tschofen et al., 2019</td>
</tr>
<tr>
<td>United States</td>
<td>2005</td>
<td>&gt;980</td>
<td>&gt;7</td>
<td>Mortality, morbidity</td>
<td>Fann et al., 2012</td>
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<tr>
<td>Canada</td>
<td>2008</td>
<td>6.7</td>
<td>0.5</td>
<td>Mortality, morbidity</td>
<td>Canadian Medical Association, 2008</td>
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<tr>
<td>Canada</td>
<td>2015</td>
<td>27</td>
<td>2</td>
<td>Mortality and morbidity; VSL</td>
<td>Smith and McDougal, 2017</td>
</tr>
</tbody>
</table>

Reduced labour productivity and other morbidity effects

40. Total annual costs of lost labour output in Canada are estimated at around €570 million (Canadian Medical Association, 2008, Smith and McDougal, 2017), constituting about 9 per cent of total economic costs of air pollution in 2008 (Canadian Medical Association, 2008).

41. In Canada and the United States of America, air pollution is calculated to result in a GDP decrease of approximately 0.1 per cent in 2060 compared to a non-polluted situation due to reduced labour productivity (OECD, 2016). Costs of morbidity per capita in North America in 2060 are projected to be around €100–€150 per year (see figure 10 below).

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Damage to crops

42. Air pollution’s effects on crops are quite significant in North American. OECD projections\(^1\) indicate that, in the United States of America, by 2060, agricultural production will fall by 4.9 per cent due to air pollution – this is the largest impact on agriculture in the entire OECD region, which, however, would not result in as large changes in GDP – the corresponding GDP decrease is projected at approximately 0.1 per cent. For Canada, the GDP decrease due to effect on crops is estimated at approximately 0.05 per cent, and damage to the agricultural sector at 0.6 per cent. Losses from reduced agricultural yields in Canada are estimated at €68 million in 2015 (Smith and McDougal, 2017).\(^{34}\)

Costs by sectors and pollutants

43. In the total costs of damage due to exposure to PM\(_{2.5}\) in the United States of America (see figure 11 below, left-hand panel), primary emissions of PM\(_{2.5}\) contribute about twice as much as contributions from secondary particles from NMVOC, NH\(_3\), SO\(_x\), or NO\(_x\) emissions. Damage costs per unit emitted pollutant are highest for PM\(_{2.5}\) (see figure 11, right-hand panel).

Figure 11

Contribution of pollutants to total damage from secondary PM\(_{2.5}\) in the United States of America in 2011 (left, Goodkind et al., 2019)\(^{10}\) and damage costs per unit emissions in the United States of America, as estimated in Tschofen et al., 2019\(^{31}\) (right).

44. About 75 per cent of total damage from air pollution in the United States of America is estimated to be caused by activities in four sectors responsible for less than 20 per cent of GDP – agriculture, energy (utilities), manufacturing industries and transport (Tschofen et al., 2019).\(^{31}\) These sectors have different pollution profiles: NH\(_3\) causes a major part of the damage for agriculture; SO\(_x\) for energy and industries; and NO\(_x\) for transport (see figure 12 below).
Intervals for sector-specific costs of these pollutants are summarized in figure 13 below – as in Europe, the highest unit damage is observed for PM$_{2.5}$, and the lowest for NMVOCs.

Figure 13
Sector-specific damage costs from air pollutants in the United States of America, thousands year 2015 euros/ton (sources – Goodkind et al., 2019$^{30}$, Schrader et al., 2018$^{35}$).

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46. Total damage costs from fossil fuel air pollution in the United States of America are estimated to €490 billion per year (Greenpeace, 2020).

E. On a global scale

Total damage and per cent of gross domestic product, reduced labour productivity and other morbidity effects

47. A recent OECD study\(^1\) estimates that, by 2060, annual costs of premature mortality worldwide will have increased from €2.4 trillion in 2015 to around €15 trillion–€20 trillion in 2060. Total damage from pain and suffering from illness is estimated to rise from €0.2 trillion in 2015 to €1.8 trillion, the annual number of lost working days is expected to reach 3.7 billion, and health-care costs €143 billion. Reduced labour productivity from air pollution is expected to cause global GDP loss of 0.4 per cent. The share of labour productivity effects on total market effects of air pollution is estimated at approximately 40 per cent (see figure 14 below). Non-market effects (costs of premature death and morbidity) exceed market effects by at least a factor of 8 (OECD, 2016).\(^1\) Given this relationship, total share of labour productivity losses in the total air pollution-related damage can be estimated at 5–6 per cent.

Figure 14
Global market costs from air pollution, copied from OECD, 2016\(^1\).

Damage to crops

48. Losses of crop yields due to air pollution vary significantly between countries – macroeconomic models indicate they range from a <1 per cent to a 22 per cent decrease, compared to “no-feedback projection”\(^1\). However, since agriculture’s share in global GDP is relatively small, the global impact of air pollution on agricultural output is not significant, corresponding to approximately 0.1 per cent of GDP reduction in 2060 (OECD, 2016).\(^1\)

Sector contributions

49. The sectors responsible for the largest contribution to global health damage costs from air pollution are road transport, household fuel combustion, agriculture and industrial coal burning (WHO and OECD, 2015).\(^13\) About 50 per cent of total health damage in OECD countries is due to pollution from road transport. In 2010, the damage cost from this sector was estimated at €690 billion (OECD, 2014).\(^17\) The global cost of air pollution from all fossil fuel combustion is estimated at approximately €7 billion per day, or 3.3 per cent of the world’s GDP. In Bulgaria, Belarus, Hungary, Romania, Serbia and Ukraine, damage costs of fossil fuel combustion are estimated to exceed 5 per cent of GDP (Greenpeace, 2020).\(^26\)

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VI. How much benefit will expected action bring in the future?

A. How large are the economic benefits that have been gained so far?

50. Retrospective assessments of economic benefits achieved from reduced air pollution are scarce; available studies indicate, however, that air quality policies in the European Union have resulted in a range of improvements during the past decades. For example, NH3 emission compliance with the European Union National Emission reduction Commitments Directive is estimated to have resulted in €14.6 billion in socioeconomic benefits from avoided premature deaths in the European Union-28 in 2016 (VSL) (Giannakis et al., 2019).38 Estimates for the Netherlands show that, in 2015, the avoided monetary health damage amounted to €35 billion per year (VOLY), compared to the “no action 1980–2015” scenario. Of this figure, 53 per cent is attributable to emission reductions in the Netherlands, while almost half is due to emission reductions in other European countries, including Belgium, France, Germany and the United Kingdom of Great Britain and Northern Ireland (Velders et al., 2020).39

51. In North America, the Clean Air Act of the United States of America was estimated to result in annual benefits of €2 trillion – this is from avoided premature deaths (VSL), morbidity, damage to crops and materials and recreational values. Of this figure, €10 billion is benefits in the agricultural sector, and €20 billion from reduced medical expenditures. The country’s GDP growth due to the health effects of Clean Air Act implementation is estimated at 0.02 per cent (United States Environmental Protection Agency (US EPA), 2011).40

B. European countries – forthcoming benefits from measures in place

52. Within Europe,41 trends for health damage from air pollution depend on the considered region. In Western and Central Europe and in the EECCA countries the damage is expected to decrease in the next decade – the total annual benefits in 2030 are estimated at €218 billion (approximately 0.9 per cent of current GDP, on average) and €17 billion (approximately 0.4 per cent of current GDP), respectively, compared to 2020 (see figure 15 below). In all, 2 per cent of the gained benefits are due to an increase in labour productivity, 6 per cent to other morbidity effects, and 92 per cent to avoided premature mortality. In South-Eastern Europe, the total damage trend is ascending – in 2030, air pollution is expected to cost €16.5 billion more than in 2020. Total avoided annual health damage from air pollution within the EMEP domain in 2030 is estimated at €219 billion (14 per cent, in relation to 2020).


41 “Europe” in the context of the present document refers to the EMEP modelling domain represented in Greenhouse Gas and Air Pollution Interactions and Synergies (GAINS) Europe (v.3). The split into subregions in section VI is the same as in section V; the only difference being that, in section VI, damage in the Eastern Europe, Caucasus and Central Asia (EECCA) countries is only estimated for the countries within the EMEP domain: Belarus, Republic of Moldova, Russian Federation (European territory) and Ukraine.
Figure 15
Estimated health benefits in European countries from agreed actions reducing air pollution (own calculations based on current legislation scenario in Amann et al., 2020).

53. Some countries are expected to face a total increase in total premature mortality by 2030 despite the actions to be taken to reduce emissions: quick population growth and ageing are factors that, in some cases, outweigh positive effects of emission reductions and improved air quality on the total health damage – this is especially pronounced for South-Eastern Europe, where changing population structure in Turkey seems to result in higher total health damage in 2030 than now (see figure 16 below).

Figure 16
Variations in the baseline damage reduction in 2030, in per cent to the 2020 level (own calculations based on current legislation scenario in Amann et al., 2020).

Abbreviations: AL, Albania; BA, Bosnia and Herzegovina; BY, Belarus; MD, Republic of Moldova; MK, North Macedonia; RU, Russian Federation; TR, Turkey; UA, Ukraine.
54. An International Institute for Applied Systems Analysis (IIASA) analysis for the European Union-27\(^4\) shows that, with already agreed measures, by 2050, the total damage from premature mortality due to exposure to PM\(_{2.5}\) is expected to decrease by 39 per cent, compared to 2020. Within the same period, deaths attributable to ground-level ozone will decline by 19 per cent.

55. Existing policy measures in the transport sector are expected to bring significant benefits in Western and Central Europe in the next decade – about €54 billion per year in 2030, compared to 2016. In this avoided damage, about 91 per cent is attributable to health effects, and 9 per cent is benefits from improved ecosystem services and prevented deterioration of buildings and materials (CE Delft, 2018).\(^24\)

C. Air pollution and climate actions – what are the co-benefits?

56. Costs of technical air pollution measures and damage from air pollution could be reduced if air pollution legislation were to be enhanced by climate and energy policies. For example, the Climate and Energy (C and E) framework adopted by the European Commission in 2014 is expected to result in reduced emissions of air pollutants of up to 10 per cent in 2030, compared to the emission levels estimated in a previously used baseline scenario that did not consider the C and E framework. When considering the C and E framework, the air pollution abatement costs become 4 per cent lower and the avoided damage costs 5 per cent higher than in the previous baseline. IIASA\(^42\) estimated that 27 per cent of the European Commission health improvement target for 2030 would be achieved through realization of the C and E framework scenario.

VII. Can damage costs be further avoided in the future?

57. Legislation in place will reduce health damage in the near future, but more benefits can be gained by raising the ambition level.

58. OECD, 2020,\(^43\) estimates that a 1 \(\mu\)g/m\(^3\) decrease in annual PM\(_{2.5}\) concentration in the European Union would increase Europe’s GDP by 0.8 per cent. A 10 per cent reduction in PM\(_{2.5}\) average concentration across Europe would increase European GDP by €93 billion–€185 billion, or €185–€370 per capita. About 95 per cent of the total effect of PM\(_{2.5}\) concentration on economic output is due to reduced labour productivity per worker.

A. European Union-27 beyond the baseline

59. Recent analysis in the second Clean Air Outlook\(^5\) shows that, within the European Union-27, implementation of measures in accordance with NAPCPs would result in additional health benefits of about €20 billion–€30 billion annually (see figure 17 below, left-hand panel). Additionally, European Union countries would gain about €400 million–€900 million annually from reduced negative effects of air pollution on crop yields (see figure 17, right-hand panel, NAPCP). These are benefits achieved without excessive costs. If all technically feasible measures were to be applied irrespective of costs (MTFR scenario), annual health benefits would reach €153 billion in 2050, accompanied by €5 billion in non-health benefits.


While baseline development will lead to a 24 per cent damage reduction in 2030, compared to 2020, introducing NAPCP measures will mean a 27 per cent reduction, and applying all technically feasible measures a 40 per cent reduction (see figure 18 below).

In the European Union transport sector, a faster uptake of zero emitting vehicles and a ban on pre-Euro 6 vehicles in all major cities would result in welfare benefits corresponding to €5.2 billion per year in 2030 due to improved health, reduced mortality (VOLY), better crop yields and biodiversity. Making transport policies even more ambitious – ban on pre-Euro 6 vehicles on all roads, road pricing, urban policies to reduce car use in cities – would bring €10.5 billion in benefits (CEDelft, 2018).

B. Potential benefits in Eastern and South-Eastern Europe

While within the European Union, air pollution reduction strategies have already brought benefits and reduced potential of additional reductions, in the Eastern and South-Eastern European countries, emission reduction potentials are much higher. For example, measures in the energy sector could result in 60 per cent reductions of SO$_2$ in relation to baseline emissions (Maas and Grennfelt, 2016). During the revision of the Protocol to Abate Acidification, Eutrophication and Ground-level Ozone in 2011, potential emission reductions
in the EECCA and non-European Union Balkan countries (MTFR scenario) in 2020 were estimated at 75 per cent lower than baseline scenario emissions for PM$_{2.5}$, and 39 per cent lower for NO$_x$, with a resulting 43 million years of life gained (Amann et al., 2011)\textsuperscript{44}

C. \textbf{More co-benefits from climate action}

63. Even greater benefits can be achieved if air pollution reduction measures are effectively combined with policies and measures targeting greenhouse gas emissions, such as fuel transitions or behavioural changes reducing energy demand. The second Clean Air Outlook\textsuperscript{5} indicates significant additional damage reductions (both health and other effects included) in the European Union-27 if air pollution reduction measures are applied in the 1.5 LIFE scenario of the European Union 2050 climate strategy vision – €76 billion for NAPCP measures, and an additional €55 billion for MTFR in 2050 (see figure 19 below).

Figure 19
\textbf{Damage, environmental and health benefits in the European Union-27 in 2050, based on Amann et al., 2020\textsuperscript{5}.}

64. Health co-benefits from different ambition levels of climate policies provided in Markandya et al., 2018\textsuperscript{45} show that, while current Nationally Determined Contributions (NDCs) would result in €200 billion lower damage in Europe (cumulative over the period 2020–2050), a 2°C target implies €800 billion–€2,100 billion health co-benefits, and with the target of 1.5°C, €2,000 billion–€2,900 billion in health damage can be avoided by climate policy. For the United States of America, health co-benefits are estimated at €500 billion with current NDCs, and up to €1,600 billion with higher ambition levels of climate policy (see figure 20 below).


Effective climate policies also bring co-benefits from crop yields. Ozone-related crop productivity improvements per capita resulting from NDCs and the 2°C target are highest in the Western part of the ECE region – in particular, the United States of America, Canada and certain European countries such as Spain (see figure 21 below).

Figure 21
Ozone-related crop yields co-benefits from climate policies, copied from Vandyck et al., 2018. Upper panel – difference between reference scenario and Nationally Determined Contributions in 2030; lower panel – difference between reference scenario and 2°C reduction scenario in 2050.

**VIII. Are the avoided costs of inaction larger than emission control costs?**

66. Several previous cost-benefit analyses supporting policy decisions in the ECE region indicated that a significant part of potential emission reductions can be done at costs that are lower than benefits gained from improved air quality. For instance, welfare benefits from the Clean Air Act of the United States of America are estimated to be more than 30 times higher than implementation costs (US EPA, 2011).40

67. According to the cost-benefit analysis (CBA) of final policy scenarios for the European Union Clean Air Package (Holland et al., 2014),18 annual net health benefits from the suggested national emission ceilings range from €42 billion to €164 billion in 2030, at costs of around €4 billion. This means that benefits are about 10–40 times higher than costs. Considering the C and E framework, which implies lower abatement costs and larger benefits (see figure 22 below), the benefit-to-cost ratio and the net benefits at the same ambition level are even higher. Furthermore, considering that the finally agreed emission reduction commitments were less ambitious than the ambition levels analysed in Holland et al. (2014), the benefit/cost ratio is higher than 10–40.

Figure 22
Costs and benefits of different ambition levels of air pollution policies (100 per cent corresponds to full implementation of all possible air pollution measures on top of the Price-Induced Market Equilibrium System 2013 reference economic scenario), in year 2005 euros. Copied from Amann et al., 201442

Abbreviations: PRIMES, Price-Induced Market Equilibrium System.

68. The latest assessment of costs and benefits from potential additional policy measures in the European Union-27 (Amann et al., 2020)5 concludes that annual net welfare benefits (VSL) from NAPCP measures in 2030 would amount to €31 billion, while the full implementation of technical measures would result in a net benefit of €146 billion. Benefits are estimated to be approximately 25 times higher than costs in the NAPCP scenario, and approximately 7 times higher than in the MTFR scenario where the costs are also high. The further analysis of macroeconomic effects also shows that total benefits are higher than abatement costs (see figure 23 below).
69. The average costs of an optimal air pollution strategy are 0.01–0.02 per cent of GDP (Maas and Grennfelt, 2016)\(^5\) – this could be compared to the approximately 5 per cent of GDP that air pollution welfare damages correspond to in Western and Central Europe.

70. Cost of action could be compared to cost of inaction also for specific pollutants, industries or facilities. Recent analysis of the steel industry in Europe (Scarborough et al., 2019)\(^47\) indicates that measures assuring Best Available Techniques conclusions compliance in the sector would bring benefits that are 3.3–14 times higher than costs. Costs of action to abate ammonia are compared to the avoided damage in the recent Task Force on Integrated Assessment Modelling Assessment report on ammonia\(^48\) – the comparison shows that benefits (€17.5/kg, as in CE Delft, 2018)\(^28\) are 1.2–4.4 times higher than costs (€4–15/kg, as in Wulf et al., 2017).\(^49\)

71. Through analysis of costs and benefits at the facility level, decision-makers could make use of the estimated damage costs per ton pollutant, available for all European ECE countries. Comparing avoided damage costs and comparing with costs of suggested technical solutions (see case study in box 3 below) could provide justification of investment decisions resulting in emission reductions beyond legally required levels.

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Box 3

Cost of action vs Cost of Inaction at the level of facilities – Case study of Apatity coal plant.

Case study of the Apatity coal plant

The Apatity combustion plant in the north-west of the Russian Federation (1,530 MWth thermal output) has been in operation since 1959, using coal as its main fuel to produce heat and power. The Expert Group on Techno-economic Issues (EGTEI) estimated annual abatement costs of installing equipment to reduce emissions of SO₂, NOₓ, and total suspended particles (TSP) – with wet flue gas desulfurisator, selective catalytic reduction, and electrostatic precipitator, respectively (see table 5 below). Costs of avoided damage to health due to these abatement techniques are estimated by applying country-specific unit damage costs as in Schucht et al., 2021, the range is €158 million–€469 million, depending on the chosen metric for health valuation. Irrespective of whether VSL or VOLY is chosen, total benefits from avoided damage significantly exceed costs. Total annual costs are estimated at €27.4 million, so the benefit-to-cost ratio lies between 6 and 17 (see figure 24 below).

Table 5

Parameters used for calculating costs and benefits of installation of cleaning technologies at Apatity coal plant, based on EGTEI, 2011, Schucht et al., 2021, and GAINS model scenarios as in Amann et al., 2020

<table>
<thead>
<tr>
<th>Pollutant</th>
<th>Emissions in 2008/2010, kt</th>
<th>Removal efficiency of equipment, per cent</th>
<th>Removed emissions, kt</th>
<th>Abatement costs, millions of year 2015 euros</th>
<th>Avoided damage, millions of year 2015 euros</th>
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<td>6.18</td>
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</tbody>
</table>

Figure 24

Costs and benefits of installation of cleaning technologies at Apatity coal plant, based on EGTEI, 2011 and Schucht et al., 2021.

IX. Closing remarks

72. Economic valuation of air pollution provides useful information on damage costs (socioeconomic welfare losses) of air pollution and thereby enables: direct comparisons of economic activities with environmental and human health effects; and valuation of reduced damages/increased benefits of further emission reductions, or, when available solutions are available but not implemented, the damage costs of inaction.

73. The results presented in this document are extracted from numerous studies carried out by several independent research groups. Although numerical values can vary due to differences in underlying assumptions, all studies show substantial welfare effects of poor air quality. Correspondingly, there is high confidence in the possibility to estimate welfare effects of changes in air quality. But it is also clear that the values in most cases are underestimations of the full welfare effects. Many health and complex ecosystem effects remain to be monetized.

74. The damage cost approach is a useful tool to assess unintentional welfare effects of new infrastructure investments or installations but requires further development. To support decisions on new projects or permits, several countries apply damage costs per unit of emission, to quickly scan potential additional damage to health and ecosystems from those activities and to decide whether additional air pollution measures are required and proportional. Often, these assessment tools only look at local or national damage, while (avoided) transboundary damage is omitted. Other important omitted damages are damages on biodiversity. A comprehensive assessment would require including all external effects, including transboundary impacts. There are also considerable information gaps between the Eastern and Western parts of the ECE region, especially regarding valuation studies carried out by East European research groups and scenarios for future air pollution levels in Eastern Europe.

75. Air pollution costs are currently substantial. For almost half of the countries in the ECE region, aggregated damage costs correspond to approximately 5 per cent of GDP. Future scenarios are foremost available for Western Europe. Through existing policies, monetary damage in the EMEP domain is be expected to be 14 per cent lower in 2030 than today. Of the European Union damage costs expected to remain in 2030, 21 per cent can be removed through additional policy actions. In most cases, the costs of reducing emissions are far lower than the corresponding reduction of damage costs; for example, the benefits of NAPCPs in the European Union-27 are more than 20 times higher than the emission control costs.

76. A summary of the methodology and data availability is presented in table 6 below, which also gives indications of comparable data sets. Major data gaps identified are:

(a) Lack of total damage estimates for recent years for ECE countries outside the EMEP domain\(^3\) (except for Canada and the United States of America) – Armenia, Azerbaijan, Israel, Kazakhstan, Kyrgyzstan, the Russian Federation (Asian part), Tajikistan, Turkmenistan and Uzbekistan – and some countries within the EMEP domain (Andorra, Iceland, Liechtenstein, Monaco, Montenegro, San Marino and Serbia);

(b) Lack of analyses of future damage in the regions outside the EMEP domain.\(^3\) \(^\text{\cite{gains}}\)

The GAINS model v.4,\(^50\) under which more countries are included, enables future assessments of damage with broader geographical coverage.

\(^{50}\) https://gains.iiasa.ac.at/models/gains_models4.html.
### Table 6

**Estimates summary of the method and main data sources**

<table>
<thead>
<tr>
<th>Aspect analysed</th>
<th>Method, main sources</th>
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| Current total monetized damage and per cent of GDP | EMEP-domain: GAINS-ARP modelling for 2020, 2030 (figures 2 and 4–5)  
Caucasus and Central Asia – mortality rates for 2010 from literature\(^{13}\) in combination with VSL (figure 3)  
North America and global – literature\(^{20,30,31,32,33,34}\), values for the period 2010–2015 (table 4) |
| Country-specific damage costs per pollutant | EECCA: literature\(^{16}\); damage only in the emitter countries (tables 1–2)  
Western and Central Europe: literature\(^{15}\); damage in EEA38+ United Kingdom (table 3), European Union-average from literature\(^{28}\) (figure 9)  
North America: average for United States only from literature\(^{31}\) (figure 11), no country-specific estimates |
| Damage costs per sector | Europe: literature\(^{10,25,26,27,28}\) (figure 8)  
North America\(^{30,35}\): literature (figure 13) |
| Labour productivity, damage to crops | Examples from literature\(^{13}\) for Western Europe (figure 6)\(^{22}\), Russian Federation, and North America (figure 10) |
| Benefits achieved in the past | Examples from literature for Europe\(^{38,39}\) and North America\(^{40}\) |
| Future benefits from measures in place | EMEP domain: GAINS-ARP modelling for 2020, 2030 (figures 15–16).  
No monetary assessments for other regions |
| Benefits beyond the baseline | Western and Central Europe: literature\(^{5}\) (figures 17–18).  
No monetary assessments for other regions |
| Potential co-benefits from climate actions | Examples from literature for Europe\(^{45}\) (figures 19–20) and North America\(^{45}\) (figure 20), a global example\(^{46}\) (figure 21) |
| Costs v. benefits | Examples from literature for Europe\(^{42,5,21,28,47,49}\) (figures 22–23) and North America\(^{40}\)  
Own CBA at plant level (box 3) based on assessments of emission reductions and damage costs from literature\(^{16}\) |

**Abbreviations:** ARP, Alpha RiskPoll.

77. It is important to continue efforts to improve the coverage of the values included in valuation studies. All the observed health effects of air pollution should be monetized, and ecosystem effects such as biodiversity effects should be more monetized. It is also important with more assessments of current and future sector-specific marginal damage costs, especially for Eastern and South-Eastern Europe.